Metal-free graphene-carbon nitride hybrids for photodegradation

of organic pollutants in water

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ABSTRACT

Hybrid photocatalysts of graphitic carbon nitride (g-C₃N₄) and reduced graphene oxide (rGO)

composites were prepared in one-pot via a thermal condensation of melamine with different

amounts of graphene oxide (GO). As metal-free hybrids, the prepared photocatalysts

presented enhanced performances in photooxidation of both methylene blue and phenol in

water solutions under various light irradiations. The level of rGO significantly affected MB

photodegradation efficiencies. The introduced graphene can improve the MB adsorption and

optical absorption in visible light region, therefore enables the hybrids to efficiently degrade

MB under visible light with wavelengths longer than 430 nm. The metal-free photocatalysts

were also able to degrade phenol effectively and the effects of catalyst loading and initial

phenol concentration were investigated. This study provided an efficient and environmentally

benign photocatalyst for degradation of organic pollutants in water, with complete prevention

of secondary contamination from metal-leaching.

Keywords: Photocatalysis; g-C₃N₄; Methylene blue; Phenol; Graphene; Metal-free

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1. Introduction

Nowadays, worldwide attention has been drawn to develop the state-of-the-art technologies to remove toxic and hazardous chemicals in various wastewaters discharged from industries and households. The most popular Fenton process generates hydroxyl radicals (•OH) to completely decompose organic compounds into water and carbon dioxide [1, 2]. However, the traditional Fenton reaction requires large amounts of chemical reagents and suffers from the generation of excess sludge and strict pH requirement (pH~3) [3, 4]. The leaching problems and the associated secondary contamination still stubbornly exist in any metal-based homogeneous and heterogeneous systems in the advanced oxidative processes (AOPs) [5, 6].

In recent years, photocatalysis has been intensively investigated as a fascinating way for oxidative reactions without consuming other oxidants [7-10]. A variety of metal-based materials were employed as efficient photocatalysts, yet the dissolved metal ions and high cost restricted their availabilities in industrial application [11, 12]. Wang and co-workers first reported that polymeric graphitic carbon nitride (g- C_3N_4) can act as a metal-free photocatalyst to motivate water splitting for hydrogen generation under visible light [13]. Generally, g- C_3N_4 can be synthesized by thermal condensation of precursors at 500 - 600 °C and it presents great thermal and chemical stabilities under ambient atmosphere and superior photoelectrochemical properties [14]. Subsequent studies indicated that g- C_3N_4 was able to be effective photocatalysts in selective oxidation [15, 16], fuel cell [17], and solar fuel production [18, 19].

Various strategies were applied to modify the structure of carbon nitride by nanocasting or soft templating to form feature-ordered pores and controlled intricate morphologies [14]. Transition metals, such as Fe [20, 21], Cu [22], Ta [23], Co [24], and Ag [25, 26] were also deposited onto g-C₃N₄ to form novel organic-inorganic composites, which strongly tuned the optical, electrical and chemical properties of g-C₃N₄, and then dramatically enhanced the photocatalysis. Yan et al. [27] reported that B-doped g-C₃N₄ could enhance the photodegradation of rhodamine B under visible light. Wang et al. [28] successfully utilized metal-free co-dopant of boron and fluorine to modify the surface chemistry and electron states of g-C₃N₄. The innovative photocatalyst presented superior performance for oxidation of clycloalkanes toward formation of cyclohexanone with notably high selectivity.

The emerging graphene, a new 2-dimentional (2D) nanocarbon material with sp² hybridized carbon lattice, has opened up a brand new field in catalysis science and technology [29, 30]. The pristine graphene possesses a large specific surface area, superb charge carriers' mobility and outstanding electrical conductivity [31]. Several studies were carried out to utilize graphene to improve the photocatalytic activity of titanium dioxide [32-34]. Both experimental results and theoretical studies revealed that the enhanced photocatalytic activity mainly derived from the chemically bonded interface between graphene and TiO₂ [33, 35]. Graphene was also applied to promote metal-free sulphur photocatalyst [36]. The graphene in the composite enhances the charge transportation, facilitates the separation of photo-excited electron-hole pairs, and enables the generated electrons transferring to graphene to coordinate redox reactions [33]. Zhang et al. [37] reported that rGO can modulate the band structure of g-C₃N₄, then significantly increase the photocurrent. Li et al. [15] reported that introduction of graphene to g-C₃N₄ can promote the conversion and selectivity in activation of O₂ for selective oxidation of secondary C-H bonds of cyclohexane. Du et al. [38] applied density functional theory to theoretically investigate the interface of graphene and g-C₃N₄, and found significant charge transfer between them, which alters the electronic properties and extends optical absorption in the visible region. Xiang et al. [39] reported that graphene/g-C₃N₄ hybrids demonstrated enhanced photocatalytic activity for hydrogen production under visible light. However, the graphene/g-C₃N₄ has not been investigated in degradation of organic pollutants in water.

More recently, we reported a novel metal-free hybrid of g-C₃N₄/carbon nanospheres for degradation of organic pollutants in water [40]. In this study, we developed an *in-situ* approach to synthesize g-C₃N₄/graphene composites via a one-pot condensation. Graphene was proven to be a promising promoter for enhancing photo-oxidation of g-C₃N₄ under visible and UV-visible light for phenol and methylene blue degradation.

2. Experimental

2.1. Synthesis of g-C₃N₄/graphene hybrids (CN-G).

Graphene oxide (GO) was prepared by a modified Hummers method [41] and detailed procedure can be found in our previous publications [5, 6]. To prepare the g- C_3N_4 /graphene hybrid, 5.0 g of melamine were mixed with fixed amount of GO (0.1, 0.2, 0.4, and 0.6 g) in 100 mL ethanol solution. The suspension was stirred and kept on a hotplate at 50 °C

overnight to dry, then collected in a crucible and transferred in a muffle furnace. The samples were annealed with a heating rate of 10 °C/min to 550 °C, kept at the temperature (unless mentioned elsewise) for 1 h, and cooled down naturally. The prepared catalysts were washed with deionized water for 3 times and denoted as CN-G-0.1, CN-G-0.2, CN-G-0.4, and CN-G-0.6 (referring to the amount of GO), respectively.

2.2. Characterization of nanocarbons

Scanning electron microscopy (SEM, Zeiss Neon 40EsB FIBSEM) was applied to investigate the morphological information of the hybrids. Transmission electron microscopy (TEM) imaging was obtained from a JEOL-2011 TEM instrument. X-ray photoelectron spectroscopy (XPS, ESCALAB 250) was carried on with an Al-K α source. X-ray diffraction (XRD) patterns were obtained on a D8-Advanecd X-ray diffractometer (Bruker, Germany) with Cu K α radiation (λ = 1.5418 Å). Micrometrics-Tristar 3000 was applied for N₂ adsorption to obtain the specific surface area (SSA), pore volume, and pore size distribution. Fourier-transform infrared (FT-IR) spectra were acquired from a PerkinElmer Spectroscopy-100. The thermal gravimetric-differential scanning calorimetry (TG-DSC) was performed on a Mettler-Toledo-Star system in air at a rate of 10 °C/min in a range of 35 - 900 °C. UV-visible diffuse reflectance spectra (UV-DRS) were acquired from a JASCO-A670 spectrophotometer. The recombination behavior of photoinduced carriers was studied by photoluminescence spectra (PL) on a Varian Cary Eclipse spectrometer at an excitation wavelength of 330 nm.

2.3. Photodegradation of phenol and methylene blue

Photocatalytic efficiencies of the CN-G samples were evaluated by degradation of phenol (20 ppm) and methylene blue (MB, 10 ppm) under various irradiations provided by a MSR 575/2 metal halide lamp (575 W, Philips) with different cut-off filters. In a typical experiment, 200 mL phenol (or MB) solution was added with 200 mg photocatalyst (1.0 g/L) into a double-jacket reactor with cycling water pumped from a water bath at 30 °C. During each interval, 1 mL of the phenol solution was withdrawn and injected into a vial though a PTFE filter (0.45 μm, 25mm). The sample was analyzed on a Varian high performance liquid chromatography (HPLC) with a UV detector at 270 nm through a C-18 column. The MB suspension solution was centrifuged to remove the solid catalysts and tested on a JASCO UV-vis spectrophotometer (664 nm).

3. Results and discussion

3.1. Physicochemical properties of the materials

The morphology of the hybrids was revealed by SEM and TEM imaging. Pristine g- C_3N_4 in Fig. 1(a) showed a slate-like bulk structure with sharp edges. Fig. 1(b) presents the typical wrinkled lamellar-like layers of reduced graphene oxide (rGO). It can be seen in Fig. 1 (c) that, after hybridization with the GO by an *in-situ* method, g- C_3N_4 formed a smaller, packed and porous structure. In Fig. 1(d), rGO sheets were observed and demonstrated a close interaction with g- C_3N_4 particles, indicating that the interfacial surfaces were formed between the rGO and g- C_3N_4 .

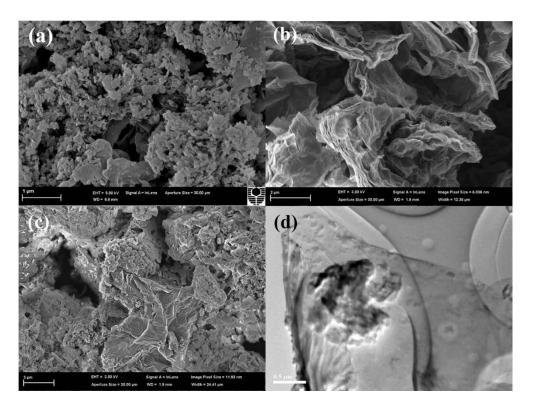


Fig. 1 SEM images of (a) g- C_3N_4 , (b) rGO-550, (c) CN-G-0.4, and (d) TEM image of CN-G-0.4.

The crystal structures of pristine g- C_3N_4 (CN) and CN/graphene composites were studied by XRD (Fig. 2). The sharp peak at 2θ of 27.5° reflects an interlayer distance of 0.324 nm, which is referred to the stacking of C-N bonds in the aromatic structure and reflects the (002) plane of graphitic materials [39]. The peak at 2θ of 13.1° can be indexed to be (001) plane (0.676 nm), related to the small tilt angularity and in-plane repeated units and reported in many references in polymerized preparation of CN with different precursors such as urea, cyanamide, and dycyanamide [8, 40, 42, 43]. No obvious peaks for GO or rGO were found in the CN-G hybrids. This can be ascribed to the fact that the graphene stacking was destroyed

during the thermal annealing and reduction process, and that the peak of rGO at around 26 $^{\circ}$ was overlapped by g-C₃N₄ [39].

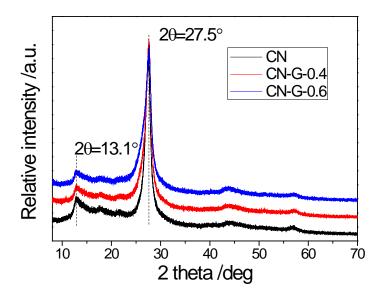


Fig. 2 XRD patterns of CN and CN-Gs.

The thermal stabilities of photocatalysts were evaluated by thermal analysis. It can be seen from TGA curves in Fig. 3(a) that g-C₃N₄ demonstrated a great stability up to 550 °C. It was well reported that the g-C₃N₄ prepared at high-temperature condensation presented a stable tri-s-triazine structure [44-46]. The rapid weight loss was attributed to breaking-up of C-N bond and combustion in air. Compared to pristine g-C₃N₄, the stability of CN-G composites decreased due to the breaking down of rGO carbon skeleton. The lower thermal stability of the chemically derived graphene was ascribed to the various defective edges and functional groups, which could act as combustion sites in air at higher temperatures. The collapse progress of the graphene network might influence the stability of CN-G hybrids. Fig 3(b) illustrates DSC curves of CN and CN-G samples. The highly-polymerized CN presented a broad endothermic peak owing to the high bonding energy of C-N bond. The exothermic peak of CN at 712 °C shifted to lower temperature for the hybrids and the intensity increased due to the decomposition of rGO.

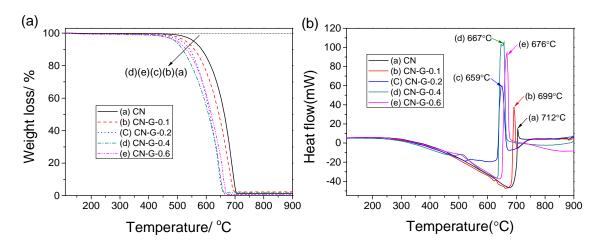


Fig. 3 TG-DSC profiles of CN and CN-G samples, (a) TGA, (b) DSC.

A comparison of FT-IR spectra between pure g-C₃N₄ and its hybrid photocatalysts were illustrated in Fig. 4. No obvious differences were observed for the CN and CN-G-0.4. The broad peak between 3000–3500 cm⁻¹ is associated with the adsorption of H₂O and CO₂ in the air [39]. Several strong bands in the region of 1200 - 1700 cm⁻¹ were found in both samples. These characteristic bands arise from the typical stretching modes of C–N heterocycles and are universally found in the g-C₃N₄ and g-C₃N₄@nanocarbon hybrids [39, 40, 47-49]. However, no C-O or C=O bonds were identified in the CN-G-0.4, suggesting that most of the oxygen functional groups on GO were removed during the thermal reduction. The obvious peak at around 800 cm⁻¹ is assigned to the typical triazine-units breathing mode [39, 50].

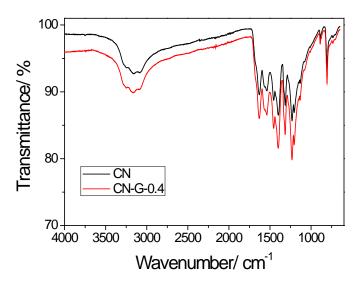


Fig.4 FT-IR spectra of CN and CN-G-0.4.

Fig. 5(a) shows XPS survey spectra of CN and CN-G-0.4. Carbon, nitrogen and oxygen contents in pure g- C_3N_4 is 48.03, 50.89 and 1.08 at.%, respectively. As the carbon nitride was thermally polymerized in the muffle furnace with static air, it was likely to form some oxygen groups on the edges of the g- C_3N_4 . The chemical composition of CN-G-0.4 was much similar to CN, suggesting that only little amount of rGO remained in the hybrids after thermal annealing. The oxygen contents increased to 1.64 at.% on CN-G-0.4 owing to the oxygen functional groups in rGO. It was suggested that phenol and quinone groups decomposed between 500 and 900 °C and the oxygen groups were completely removed at ~1100 °C [51].

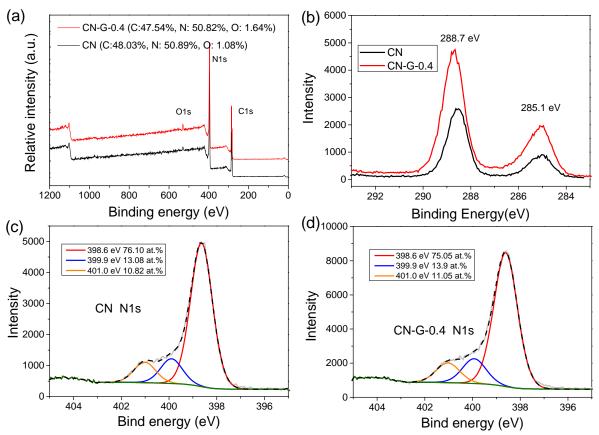


Fig. 5 (a) XPS survey of CN and CN-G-0.4, (b) C1s scan of CN and CN-G-0, (c) N1s high resolution scan of CN and (d) N1s high resolution of CN-G-0.4

Fig. 5(b) shows that the C1s peaks consist of two parts. The peak at the bonding energy of 285.1 eV is corresponding to the sp^2 C-C bond and the other sharp peak at 288.7 eV is attributed to the carbon adjacent with three N atoms [39]. High resolution N1s XPS was performed to probe the N categories on the catalysts. The N1s peak (Figs. 5(c) and (d)) can be fitted into three parts located at 401.0, 399.9, and 398.6 eV. The peak at 398.6 eV is associated with the sp^2 hybridized N bonded with two carbon atoms (C-N=C), which is the dominant nitrogen species in the g-C₃N₄ with tri-s-triazine structure [52]. The other two

peaks, 401.0 and 399.9 eV, could be assigned to the quaternary N (N bonded with three carbon atoms, N-(C)₃) and amino groups (C-N-H), respectively [39, 52, 53].

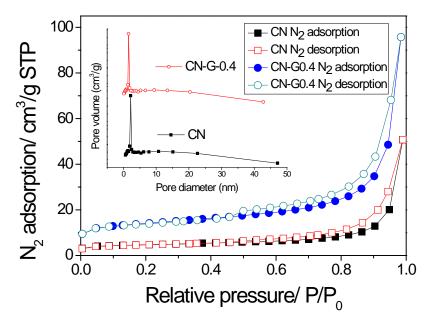


Fig. 6 N₂ sorption and pore size distribution of CN and CN-G-0.4.

The pore structures and specific surface area (SSA) of CN and CN-Gs were investigated by N_2 adsorption and desorption isotherms. Fig. 6 indicates mesoporous structure of the carbon catalysts (2 - 50 nm). The CN/graphene exhibits higher pore volume (0.118 vs 0.067 cm³/g of CN) and larger SSA (44.63 vs 14.73 m²/g of CN). But the porous hybrids present a smaller pore size.

Fig. 7 shows UV-vis DRS of CN and CN-G samples. The pristine g-C₃N₄ presents an absorption edge at 465 nm associated with a band gap energy of 2.67 eV, which is in good agreement with reported value [40]. As shown in the inset, pristine g-C₃N₄ was bright yellow, and the introduction of rGO to CN would make the color changed to brown, dark blue and black at increasing rGO loadings. Accordingly, the red-shift derived from hybridization of rGO was observed. The absorption thresholds of CN-G-0.1, CN-G-0.2, CN-G-0.4 and CN-G-0.6 were 494.0, 494.2, 509.8 and 533.0 nm, respectively, and the band gap energies were then evaluated to be 2.51, 2.51, 2.43 and 2.33 eV, respectively. The results suggested that graphene can effectively increase the absorption in visible light region.

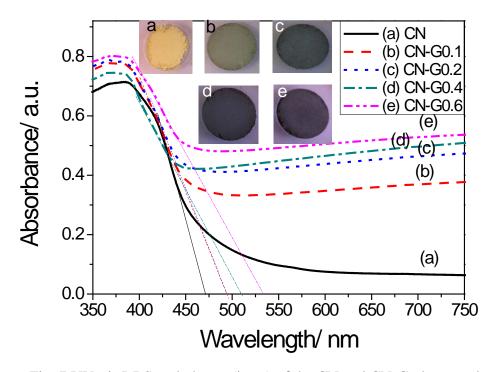
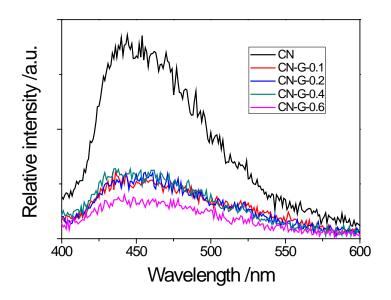


Fig. 7 UV-vis DRS and photos (inset) of the CN and CN-G photocatalysts.

Apart from the light absorption, the photodegradation efficiency is also controlled by the separation rate of the photoinduced carriers [40, 54]. Fig. 8 shows the PL spectra of pristine and modified g-C₃N₄ samples. The pristine g-C₃N₄ showed the highest intensity of photoluminescence, indicating the strongest recombination of carriers and the lowest photocatalytic activity. CN-G-0.1, 0.2 and 0.4 samples showed similar carries recombination, and CN-G-0.6 presented the lowest intensity.



3.2. Photocatalytic oxidation of MB and phenol

Fig. 9 (a) shows control experiments of MB and phenol adsorption on G-CN-0.4 and photolysis of MB without a catalyst. MB could hardly be degraded under UV-visible light without a catalyst. The adsorption tests indicated that about 27% of MB and 4% of phenol were removed due to adsorption on CN-G-0.4. MB is a cationic dye and tends to be adsorbed on the porous materials, especially on the nanocarbons with oxygen functional groups.

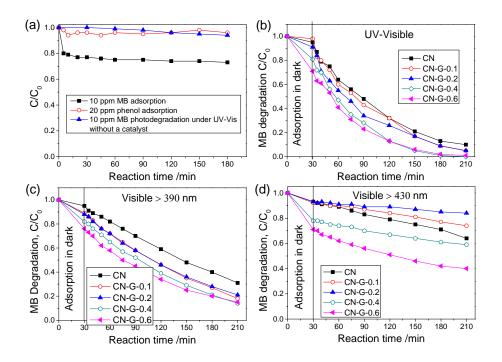


Fig. 9 Control experiments of adsorption on CN-G-0.4 and photolysis of MB under UV-visible light without a catalyst (a), and photodegradation of MB under various conditions: UV-visible light (b), visible light > 390 nm (c), and (d) visible light > 430 nm. [Catalyst = 0.5 g/L; MB = 10 ppm; Temperature = 30 °C].

Fig. 9 (b) shows the effect of rGO contents on adsorption and photodegradation of MB under UV-vis irradiations. Pristine g-C₃N₄ showed minor adsorption of MB and only 3% MB was adsorbed in 30 min. Introduction of rGO significantly improved the MB adsorption of CN-G samples, and higher content of rGO led to higher adsorption. About 20% and 30% of MB were removed in 30 min by adsorption on CN-G-0.4 and CN-G-0.6, respectively. Around 90% of MB was decomposed on pristine g-C₃N₄ by adsorption and 180 min photodegradation. The MB removal increased to 95% on CN-G-0.1 and CN-G-0.2 and 100% MB removal was achieved on CN-G-0.4 and CN-G-0.6. Fig. 9 (c) shows visible light photodegradation of MB

under light of $\lambda > 390$ nm. Roughly, 69% of MB was removed by g-C₃N₄ and 82%, 79%, 86%, and 85% MB removals were achieved on CN-G-0.1, -0.2, -0.4, and -0.6 accordingly. Compared with the pristine g-C₃N₄, enhanced MB photodegradation under visible light was observed on CN/graphene hybrids. Fig. 9 (d) shows the activities of the CN-G samples under visible light with wavelength longer than 430 nm. Approximate 36% MB decomposition was achieved on g-C₃N₄ in 210 min, whereas only 26% and 16% of MB were removed by CN-G-0.1 and CN-G-0.2, respectively. Meanwhile, CN-G-0.4 and CN-G-0.6 provided 41% and 60% MB degradation, respectively. It was reported that increasing graphene contents could enhance both MB adsorption and visible light absorption, and then possibly improve the photocatalytic activity [38]. It can been seen in Fig 9(d) that the improved MB removal efficiency under visible light (> 430 nm) was mainly due to the enhanced MB adsorption with high graphene contents owing to the strong π - π interaction between MB and graphene network. When the graphene content is low, MB adsorption and optical absorption cannot be significantly improved. The complicated processes involving adsorption, light absorbance, microstructure and surface feature led to a lower MB removal on CN-G-0.1 and CN-G-0.2 than pristine CN.

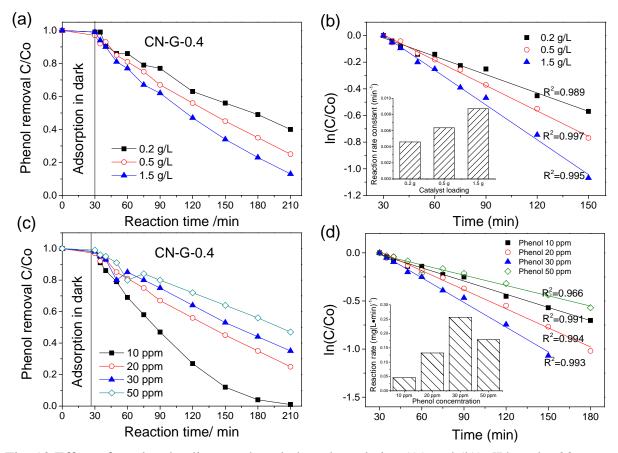


Fig. 10 Effect of catalyst loading on phenol photodegradation ((a) and (b)), [Phenol = 20 ppm;

Temperature = 30 °C; Irradiation: UV-visible light], and effect of initial phenol concentration on phenol photodegradation (c and d). [Catalyst = 0.5 g/L; Temperature = 30 °C; Irradiation: UV-visible light].

CN-G-0.4 was further applied to photodegrade phenol under UV-visible light. Fig. 10(a) indicates that phenol could barely be removed by adsorption on CN-G-0.4. However, catalyst loading poses significant influence on phenol removal. Around 60% phenol removal was achieved in 210 min at a catalyst dosage of 0.2 g/L. While, 75% and 87% of phenol were degraded when the catalyst loading was increased to 0.5 and 1.5 g/L, respectively. The apparent reaction rate constants (Fig. 10(b)) under different catalyst loading of 0.2, 0.5, and 1.5 g/L were estimated to be 4.59×10^{-3} , 6.35×10^{-3} , and 8.71×10^{-3} min⁻¹, respectively. Increased catalyst loading produced more photo-generated active sites, which dramatically enhanced the phenol degradation efficiency. The effect of initial phenol concentration was investigated and shown in Fig. 10(c). The phenol removal efficiency decreased with increasing initial phenol concentration. Phenol could be completely decomposed in 210 min under a low initial concentration (10 ppm), whereas 75%, 65% and 53% phenol were removed when initial concentration was increased to 20, 30, and 50 ppm. The reaction rates with initial phenol concentrations of 10, 20, 30 and 50 ppm were calculated to be 0.046, 0.132, 0.257 and 0.178 mg (L \times min)⁻¹, respectively. Fig. 10(d) shows that the reaction rates climbed up with increased initial phenol levels from 10 to 30 ppm and dropped when the initial phenol concentration reached 50 ppm. Increasing initial phenol levels in a proper range was able to increase the reaction rate as the phenol was quickly decomposed by the photoproduced holes and prevented the recombination. However, the excess phenol in aqueous solution would be adsorbed on the surface of the catalysts and weaken the photo-activation processes. The finite active sites generated from the fixed amount of photocatalyst could be another limiting factor as well.

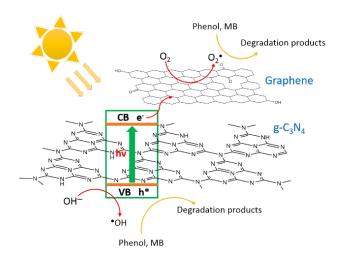


Fig. 11 Photodegradation of organics on g-C₃N₄/graphene hybrids.

3.3. Roles of graphene for enhanced photocatalysis

Pure g-C₃N₄ is more sensitive to UV and severely suffered from the drawbacks of quick recombination of the photogenerated electron/hole pairs [38]. Ge et al. [52] introduced multiwalled carbon nanotubes (MWCNTs) as a channel to simulate the photogenerated electrons transferring from g-C₃N₄ surface to MWCNTs, which presented a lower Fermi level than g-C₃N₄. The g-C₃N₄/MWCNTs hybrids presented a higher photocurrent intensity and a better photocatalytic activity toward hydrogen evolution. Sun and co-workers [40] synthesized g-C₃N₄/carbon sphere (CS) composites via a hydrothermal method. The hydrothermal process was able to modify the crystal structure of g-C₃N₄ and improve phenol adsorption. The interfaces between g-C₃N₄ and CS effectively prevented the recombination of electron/hole pairs and enhance light absorption under visible light region. Graphene possesses a similar layer structure with g-C₃N₄ and optical chemical and electronic properties [15, 31, 55]. The mechanism of graphene promoted photodegradation of MB and phenol was illustrated in Fig. 11. Introduced graphene could significantly improve the optical absorption under visible light and increased the photo-oxidative degradation as shown in Figs. 9 and 10. The CN/graphene hybrids facilitated the excited electrons from the valence band of g-C₃N₄ to the conduction band followed by a fast transportation to the graphene sheet, which results in a lower recombination rate between the photo-motivated holes and electrons [38, 39]. The holes would abstract one electron from the target organic compound (MB or phenol) and the chemically unstable organic substance was quickly transferred to degradation products. Besides, the extra electron, which was transported into the sp² hybridized carbon honeycomb, might be restricted at the Lewis basic sites (quinone and ketonic groups, -C=O) of the

reduced graphene oxide but can transfer to the water and oxygen molecules from graphene basal plane and the defective sites (vacancy and edges) to form active radicals for attacking MB and phenol to degradation products [15, 40]. Additionally, it was well reported that carbonaceous intermediates could improve phenol and visible light absorption through dyesensitization, then enhancing the photooxidation process [40, 56].

4. Conclusions

CN-graphene composites were prepared by a simple thermal condensation of melamine with GO at 550 °C. Compared to pristine carbon nitride, enhanced photocatalytic oxidation of MB and phenol under visible and UV-visible light was achieved. The introduced graphene was able to modulate both the adsorption capacity and optical absorption in visible light radiation. Increased level of rGO can significantly improve the adsorption of MB on CN-G samples, extend the visible light absorption, and then enhance the visible light photodegradation of MB. The hybrid photocatalysts were also able to efficiently degrade phenol solutions. Both catalyst loading and initial phenol concentration showed significant effects on phenol removal under UV-visible light. This study provided an efficient and green photocatalyst for degradation of dye and other organic pollutants, and can contribute to the development of sustainable remediation technologies.

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