



cambridge.org/enc

## Subject Review

**Cite this article:** Quenta-Herrera E et al. (2022). Mountain freshwater ecosystems and protected areas in the tropical Andes: insights and gaps for climate change adaptation. *Environmental Conservation* 49: 17–26. doi: [10.1017/S0376892921000382](https://doi.org/10.1017/S0376892921000382)

Received: 8 February 2021  
Revised: 10 October 2021  
Accepted: 11 October 2021  
First published online: 2 December 2021


### Keywords:

biodiversity; freshwater; glaciers; mountains; protected areas; tropical Andes

### Author for correspondence:

Dr Estefania Quenta-Herrera,  
Email: [estefaniaqh@gmail.com](mailto:estefaniaqh@gmail.com)

# Mountain freshwater ecosystems and protected areas in the tropical Andes: insights and gaps for climate change adaptation

Estefania Quenta-Herrera<sup>1</sup> , Verónica Crespo-Pérez<sup>2</sup>, Bryan G Mark<sup>3</sup>, Ana Lía Gonzales<sup>4</sup> and Aino Kulonen<sup>5</sup>

<sup>1</sup>Unidad de Calidad Ambiental, Instituto de Ecología, Universidad Mayor San Andrés, La Paz, Bolivia, calle 27 Cota Cota, La Paz, Bolivia; <sup>2</sup>Laboratorio de Limnología, Museo de Zoología QCAZ I, Escuela de Ciencias Biológicas, Pontificia Universidad Católica del Ecuador, Av. 12 de octubre 1076 y Roca, Apartado: 17-01-2184, Quito, Ecuador; <sup>3</sup>Department of Geography, The Ohio State University, Columbus, OH, USA; <sup>4</sup>Consorcio para el Desarrollo Sostenible de la Ecorregión Andina (CONDESAN), Proyecto Adaptación a los impactos del cambio climático en recursos hídricos de los Andes, AICCA, La Paz, Bolivia and <sup>5</sup>Mountain Research Initiative, Mittelstrasse 43, CH-3012, Centre for Development and Environment, University of Bern, Bern, Switzerland

## Summary

Although protected areas (PAs) play an important role in ecosystem conservation and climate change adaptation, no systematic information is available on PA protection of high-elevation freshwater ecosystems (e.g., lakes and watersheds with glaciers), their biodiversity and their ecosystem services in the tropical Andes. We therefore combined a literature review and map analysis of PAs of International Union for Conservation of Nature (IUCN) and national systems of PAs and freshwater ecosystems. We found that seven national parks were created for water resources protection but were not designed for freshwater conservation (i.e., larger watersheds). High-value biodiversity sites have not been protected, and new local PAs were created due to water resource needs. We quantified 31 Ramsar sites and observed that PAs cover 12% of lakes, 31% of glacial lakes and 12% of the total stream length in the tropical Andes. Additionally, 120 watersheds (average area 631 km<sup>2</sup>) with glaciers and 40% of the total glacier surface area were covered by PAs. Future research into the role of PAs in ecosystem services provision and more detailed freshwater inventories within and around PAs, especially for those dependent on glacier runoff, will fill key knowledge gaps for freshwater conservation and climate change adaptation in the tropical Andes.

## Introduction

Protected areas (PAs) are crucial for ecosystem conservation and resilience to climate-related impacts. Reducing carbon dioxide emissions, building carbon stocks, buffering environmental changes (Walker et al. 2009, Jantz et al. 2014) and maintaining ecosystem productivity for water and food provision are among the main contributions of PAs to climate change adaptation (Dudley et al. 2010). Furthermore, Chapter 3 of the Intergovernmental Panel on Climate Change (IPCC) fifth assessment report highlighted the importance of protecting and restoring freshwater habitats as adaptive measures that benefit climate change adaptation planning (Jiménez Cisneros et al. 2014). Nevertheless, PAs such as national parks have been based principally on the needs of terrestrial ecosystems and much less on those of freshwater ecosystems (Azevedo-Santos et al. 2019, Acreman et al. 2020). The absence of a watershed approach, the lack of large rivers representation and the absence of river connectivity protection for migratory species are among the main reasons for the weak effectiveness of freshwater ecosystem conservation by PAs (Acreman et al. 2020). Although the Convention on Biological Diversity (CBD) included a conservation target of 17% of terrestrial and inland water areas in the Aichi Biodiversity Targets (Convention on Biological Diversity 2010), at the global scale, c. 70% of river reaches have no PAs in their upstream catchments, only 11% achieve full integrated protection and the average surface area of protection of the largest basins remains below 10% (Abell et al. 2017). Furthermore, other actions such as the Convention on Wetlands have emerged due to the need for wetland conservation; currently, there are 2404 wetlands (total area 2 543 391.59 km<sup>2</sup>) with Ramsar designation worldwide (Ramsar 2020).

Approximately 40% of mountain ranges worldwide do not contain any PAs, and 75% of mountain ranges have less than half of their elevational gradients covered by the strict International Union for Conservation of Nature (IUCN) categories of PAs (Elsen et al. 2018). This highlights the need to increase protection across elevational gradients, especially in sites where high impacts of warming are predicted. The tropical Andes are highly sensitive to climate change, as it is predicted that temperature will increase by more than 4°C in the

© The Author(s), 2021. Published by Cambridge University Press on behalf of Foundation for Environmental Conservation.

**CAMBRIDGE**  
UNIVERSITY PRESS



tropics, especially at high-elevation sites, due to their geographical position and steep elevational gradients (Bradley et al. 2006, Pepin et al. 2015). Increasing warming has negative impacts on freshwater resources. For instance, rapid mountain glacier retreat and changes in precipitation patterns have resulted in hydrological changes – droughts in some places and floods in others – affecting millions of people in the tropical Andes (Baraer et al. 2012, Satgé et al. 2017, Vuille et al. 2018). Furthermore, species distribution shifts to upper elevations of mountains due to increasing warming and/or new habitats created by glacier retreat (Seimon et al. 2017, Zimmer et al. 2018) call for ecosystem protection across different elevational gradients. Therefore, assessing the role of PAs in freshwater ecosystem conservation in the tropical Andes will provide an important benchmark and offer insights into climate change adaptation and remaining gaps in understanding.

We address this issue by reviewing the literature and analysing maps related to freshwater ecosystems covered by PAs in the tropical Andes. We have three research questions: (1) Do the objectives of PAs consider freshwater ecosystems conservation in the tropical Andes? (2) To what extent are mountain freshwater ecosystems protected by PAs in the tropical Andes? (3) To what extent are ecosystem services and biodiversity safeguarded by current protection efforts? The meaning of ‘protection’ can be complex, and for freshwater ecosystems protection can include a variety of actions, such as establishing new PAs and sustainable management measures to maintain river connectivity or water quality, among others. Here, the meaning of ‘protection’ is based on the PA definition of the IUCN: a PA is a clearly defined geographical space, recognized, dedicated and managed, through legal or other effective means, to achieve the long-term conservation of nature with associated ecosystem services and cultural values (Dudley et al. 2013).

## Methods

We conducted a literature review and assembled novel databases to analyse and produce maps of freshwater protection by PAs (including lakes, streams and watersheds with glaciers). For the literature review, we applied a systematic method by employing eligibility criteria and a protocol to strengthen the transparency, accuracy and completeness of reports (Moher et al. 2009), as in some other environmental studies (McDowell et al. 2019, Acreman et al. 2020). The map analysis drew on databases of leading institutions: the IUCN, the World Wildlife Fund (WWF), the Critical Ecosystem Partnership Fund (CEPF) and the National Snow and Ice Data Center (NSIDC).

### Literature review

First, we designed a protocol considering our research questions, spatial and temporal scopes and languages. The study was limited to the tropical Andes hotspot, a defined area with exceptional concentrations of endemic species and habitat loss, whose boundaries have been determined by biological commonalities (Myers et al. 2000). For question (1), related to the objectives of the creation of PAs, we reviewed literature published since 1990. For questions (2) and (3), we reviewed literature published since 2014. We considered mainly articles written in English obtained from Web of Science and Scopus, as well as research articles and reports from non-governmental organizations (NGOs) and governmental institutions in Spanish (e.g., national systems of PAs) available online. The criteria for choosing literature items

were: research articles and reports of leading institutions; documents that answered each research question in part or in full (e.g., papers that assessed the role of PAs in freshwater protection directly and/or papers that quantified freshwater ecosystems within a PA); and information on current issues regarding freshwater ecosystems and biodiversity conservation, such as climate change and habitat disruption, which highlighted the importance of the role of PAs in addressing these issues.

We conducted the literature review in a stepwise fashion. First, we produced 24 string codes with keywords on the topic for the literature search (Supplementary Table S1, available online), resulting in 1341 selected articles. Second, we verified the criteria in the titles, then in the abstracts and finally in the full documents, resulting in 359, 269 and 130 articles without duplicates, respectively. Third, we summarized each document in order to separate the information directly or indirectly related to the topic, resulting in 32 research articles that were finally selected. In addition, we included grey literature from Google Scholar and governmental institutions and NGOs in order to better answer the research questions. The grey literature review resulted in 15 documents being selected (13 management plans of PAs and 2 reports). We cross-checked from selected research articles to ensure that relevant research articles were considered in the analysis. This resulted in the inclusion of 5 research papers, leading to a grand total of 52 documents (37 research articles, 13 management plans and 2 NGO reports; Fig. S1).

### Maps of freshwater ecosystems and PAs

We overlaid PA coverages on freshwater systems layers (i.e., lakes, glacial lakes, streams and watersheds) to quantify freshwater ecosystem protection using Geographic Information System Tools. The tropical Andes hotspot layer was obtained from the CEPF (Hoffman et al. 2016). The PA information was obtained from the World Database on Protected Areas (WDPA; <https://www.protectedplanet.net>), which includes shapefiles and a robust attribute tables for each PA with IUCN categories and national levels of protection. Shapefiles of freshwater ecosystems (i.e., lakes, rivers and watersheds) were obtained from the Hydrological data and maps based on Shuttle Elevation Derivatives at multiple Scales (HydroSHEDS; Lehner et al. 2008; <http://www.hydrosheds.org>). The HydroLAKES database provides information on global lakes and reservoirs and was designed as a digital map repository to include all lakes with a surface area of at least 0.1 km<sup>2</sup> (Messenger et al. 2016). In our study, we considered only lakes for analysing lake protection. Layers of glaciers and glacial lakes came from the NSIDC: Global Land Ice Measurements from Space (GLIMS & NSIDC 2018) and the High Mountain Asia Near-Global Multi-Decadal Glacial Lake Inventory, Version 1 (Shugar et al. 2020). We selected glacier coverages of the Randolph Glacier Inventory from the GLIMS database. Furthermore, we verified Argentinean glaciers from the Argentinian Glacier Inventory (Zalazar et al. 2020) and the Atlas of Argentinean Glaciers (<http://www.glaciaresargentinos.gov.ar>), which are also available in the GLIMS database. We also quantified Ramsar sites (data from <https://rsis.ramsar.org>), which are PAs that receive the Ramsar designation after meeting criteria such as sites containing representative, rare or unique wetland types. They are compatible with all IUCN categories of PAs (Dudley et al. 2013).

Before quantifying the freshwater ecosystem protection, we selected PAs based on the IUCN classification because it facilitates comparisons among countries (Dudley et al. 2013). Definitions of each IUCN category are available in Table S4. The preliminary

assessment of the maps showed a total of 1330 PAs that cover 33% of the total surface area of the tropical Andes (1 550 539 km<sup>2</sup>). Of the total area, 0.01%, 8.20%, 0.33%, 0.58%, 3.18% and 5.03% were protected by the IUCN categories Ia, II, III, IV, V and VI, respectively, while 4.24%, 0.08% and 11.15% of the total area were under no category of protection (the not applicable, assigned and not reported categories, respectively; Fig. S2). For our study, we selected only IUCN categories from Ia to IV, as they represent the highest levels of protection from national governments. All PAs from Ecuador and Bolivia were classified as ‘Not Reported’ (149 PAs; Fig. S2c), meaning that either the IUCN category was unknown and/or the data provider had not provided any related information. Thus, for these countries, we selected PAs under the national systems of PAs: *Servicio Nacional de Áreas Protegidas* (SERNAP) for Bolivia and the *Sistema Nacional de Áreas Protegidas* (SNAP) for Ecuador (Fig. S3). Then, protected freshwater ecosystems were quantified by overlapping the layers of selected PAs onto lakes and streams. We analysed stream orders using the Strahler stream order to provide an indication of the stream-type protection (i.e., ephemeral, perennial or intermittent), as well as about how much of the headwater and lowland streams had been covered by PAs (Hansen 2001, Biggs et al. 2017). Moreover, for quantifying the number and surface area of glacier system protection (i.e., lakes, watersheds with glaciers and glaciers), we overlaid PA layers onto the layers of glaciers, glacial lakes and watersheds. For quantifying the watershed protection covered by glaciers, we calculated the number of watersheds covered by glaciers and PAs.

## Results

Of the 52 documents selected for the literature review, 44.2% (13 management plans and 10 research articles) were related to the inclusion of water resources conservation in the objectives of the PAs’ creation and related to the creation of local PAs. Moreover, 25.0% (13) and 17.3% (9) of the documents were related to freshwater ecosystem inventories and aquatic biodiversity in PAs and climate-related impacts, respectively. Furthermore, 13.5% (7) of the documents were related to ecosystem services in PAs (Table S2).

### *Do the objectives of PAs consider freshwater ecosystem conservation in the tropical Andes?*

#### *Creation of PAs and water resources conservation*

Most PAs worldwide were not created for freshwater conservation (Dudley et al. 2013). In the Andean region, Messerli et al. (1997) published the first study about the importance of including water resources protection in policy to avoid conflicts related to water scarcity. They presented a map with 60 PAs and suggested that more water PAs should be created, especially in arid places with less than 500 mm of annual precipitation. In the grey literature, we found 7 national parks out of the 16 assessed with biodiversity and water resources conservation among the main objectives of their creation (Table S3). For instance, Tunari National Park in Bolivia was created for hydrological protection and forest conservation (Table S3). In Ecuador, Podocarpus National Park was created for forest and freshwater ecosystem protection (Table S3). In the same country, the Pro-Cuencas Podocarpus Fund aimed to conserve water quantity and quality by reducing pollution by 25% and reforesting 10% of the watersheds (Redondo-Brenes 2009). In Peru, the Río Abiseo National Park

was created to protect humid forest protection and to maintain the hydrological stability of catchments (Young 1993). In Chile, Lauca National Park covers three volcanic cones, lakes and peatlands. This park and its neighbours, Las Vicuñas National Reserve and the National Monument Salar de Surire, form the Lauca Biosphere Reserve (Rundel & Palma 2000). Furthermore, we found four PAs with unclear objectives and/or that were not created for freshwater protection but the management plans highlighted the need for water resources protection. Additionally, we found five PAs with high values for glacier and freshwater conservation (Table S3). Tuni Condoriri National Park in Bolivia was the most relevant case because no management plan has yet been initiated, despite the fact that the park includes high-value glaciated mountains that provide drinking water to millions of people in the cities of La Paz and El Alto (Hoffmann & Oetting 2011).

#### *Creation of local PAs*

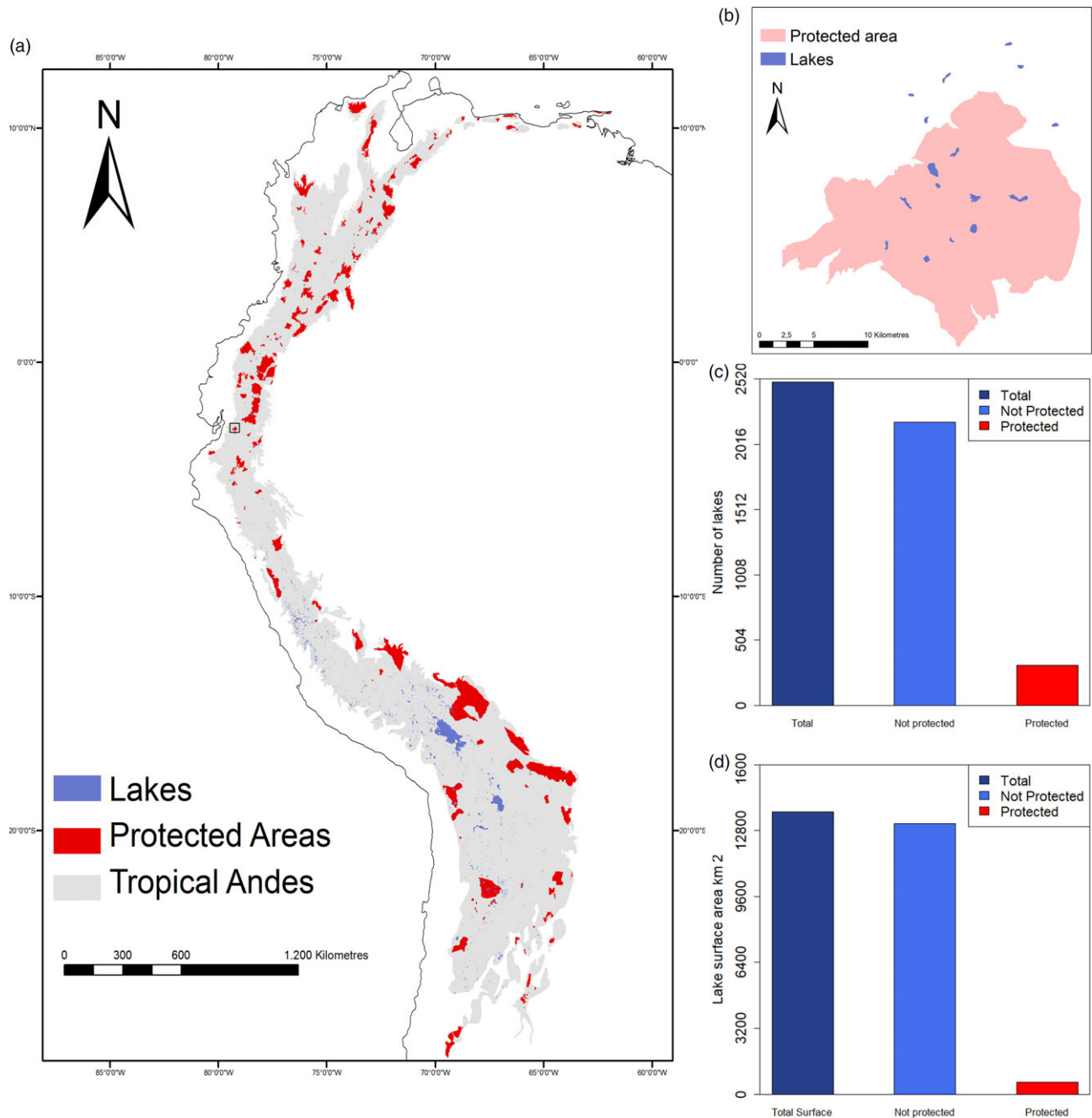
We found six studies on the creation of local PAs for water resources. In Peru, the regional Government of San Martín created the Rumialba Ecological Recovery and Conservation Zone (ZOCRE) of 23.96 km<sup>2</sup> covering the upper watersheds of the Rumiayacu, Mishquiyacu and Almendra rivers, which provide drinking water to c. 50 000 people (Montoya-Zumaeta et al. 2019). In the same country, stakeholders of the Piura community agreed to a programme to protect a large watershed spanning from the Andes to the Pacific coast through the Regional Fund for Water and Sanitation (FORASAN). Among their main activities were the establishment of private PAs such as Cuyas and Samanga (Ostovar 2019). In Ecuador, Iñiguez Gallardo et al. (2013) assessed the governance of a proposed Ramsar wetland (Saraguro–Oña–Yacuambi) containing several lakes (e.g., Laguna Grande and Tres Lagunas) and located within the PA of Shincata Protected Forest and the municipal Yacuambi Natural Reserve. In Ecuador as well, payment for ecosystem services has been one of the ways by which to address water resources protection. For instance, citizens and NGOs initiated payment to rural communities of the Antisana and Cayambe–Coca ecological reserves to guarantee water provision, estimated to be c. 80% of the supply for Quito (Joslin 2020). In Colombia and Venezuela, Leroy (2019) assessed the perceptions of two local communities regarding water scarcity and climate change indicators (e.g., disturbance of the seasons); Venezuelan farmers adopted irrigation systems and actions to protect wetlands by establishing protection boundaries, especially after a severe drought occurred.

### *To what extent are mountain freshwater ecosystems protected by PAs in the tropical Andes?*

#### *PAs, lake and glacial lake inventories*

We found that 4% (581 km<sup>2</sup>) of the total lake area (13 716 km<sup>2</sup>), representing a volume of 4.4 km<sup>3</sup>, was covered by PAs in the tropical Andes (Fig. 1a–c), and the majority of lakes (2188 lakes, 88%) were not covered by PAs (Fig. 1d). Additionally, 31% (45) of glacial lakes were covered by PAs (Fig. 3d), representing 29% (15 km<sup>2</sup>) of the total surface area of glacial lakes.

In our literature review, we found two national park inventories of lake areas. A more detailed inventory reported 202 lakes and 5955 water bodies in the Cajas National Park, Ecuador. Lakes larger than 104 m<sup>2</sup> constituted 85% of superficial water in the park, and 10 lakes deeper than 18 m contained 50% of the water sources (Mosquera et al. 2017). Similarly, Polk et al. (2017) reported lakes



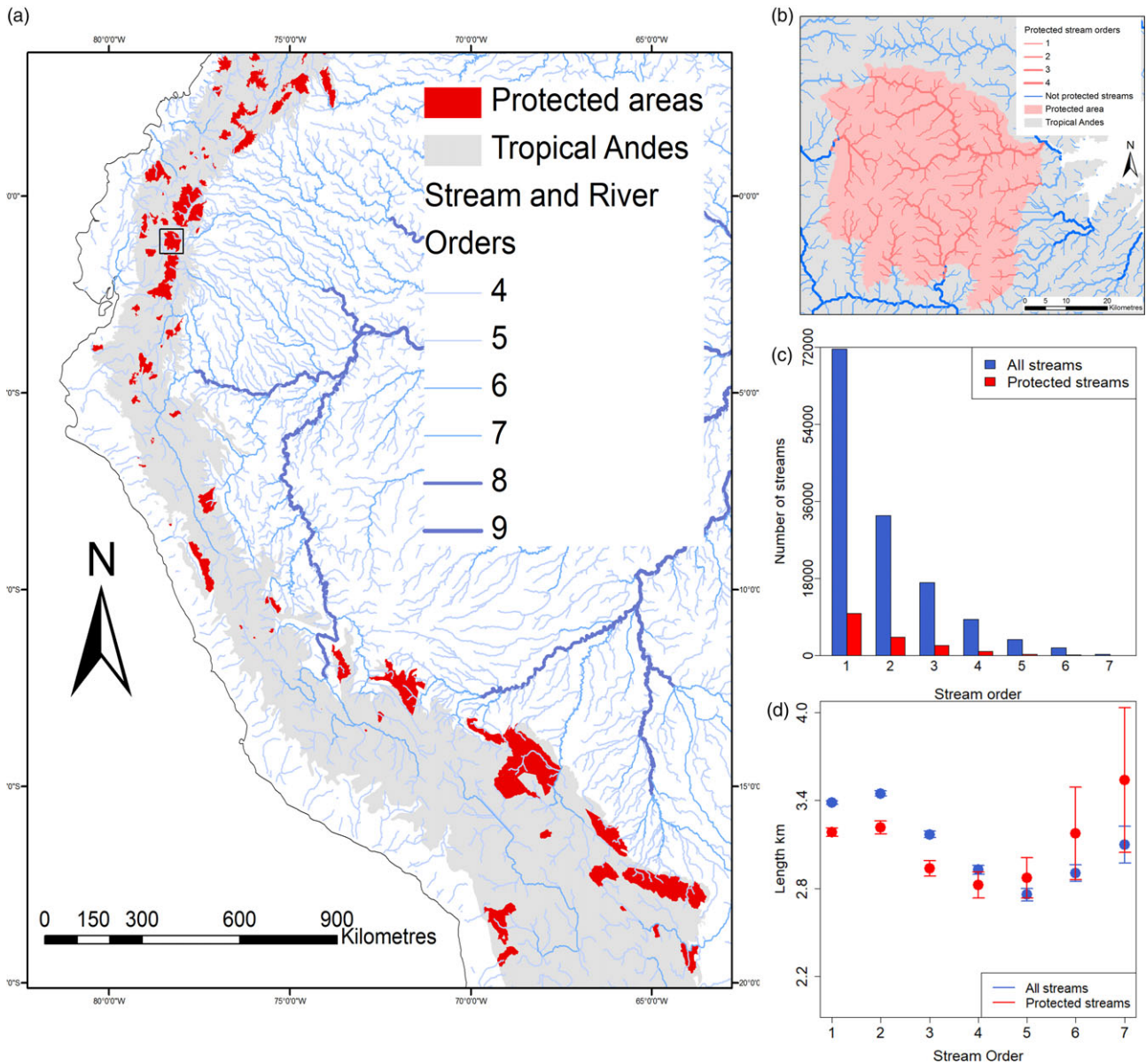
**Fig. 1.** (a) Protected areas (PAs) and lakes in the tropical Andes. The PA in the black rectangle is Cajas National Park, which is depicted in (b) for visualization of the lake distribution inside of the park. (c) Total number of PA protected and unprotected lakes. (d) Total surface area of PA protected and unprotected lakes.

ranging in area from 0.66 to 16.29 km<sup>2</sup> between 1987 and 1995 in the Huascarán National Park, Peru.

We did not find studies on the role of PAs in addressing climate-related impacts on freshwater ecosystems, only studies related to water losses by evaporation and the increase in the number of lakes due to glacier retreat in the tropical Andes. For instance, higher evaporation rates (1700 mm year<sup>-1</sup>) over Lake Titicaca were reported (Pillco Zolá et al. 2019). Moreover, 201 sites might become lakes in the future due to glacier retreat in Peruvian cordilleras (Colonia et al. 2017). Similarly, glacier recession

between 1986 and 2014 in the Bolivian cordilleras increased the number of proglacial lakes by 47% in the Cordillera Real (from 92 to 135 lakes) and by 72% in the Cordillera Apolobamba (from 29 to 50 lakes; Cook et al. 2016). Similarly, glacier retreat in the Peruvian Vilcanota–Urubamba basins augmented the lake area and number from 23.3 km<sup>2</sup> (460 lakes) in 1988 to 26.9 km<sup>2</sup> (544 lakes) in 2016, and future lake areas could grow by between 3.2% and 6.0% under IPCC scenarios Representative Concentration Pathway (RCP) 2.6 and RCP 8.5, respectively (Drenkhan et al. 2018).





**Fig. 2.** (a) Protected areas (PAs) and streams in the tropical Andes. Only a portion of the tropical Andes is shown, and only streams of fourth order and higher are depicted for better visualization. Stream orders are based on the Strahler stream order. The PA within the black rectangle is the Llanganates National Park, which is depicted in (b). (c) Total number of PA protected streams sorted by stream order. (d) Stream length in relation to stream order. Error bars show standard errors.

*PAs, streams and watersheds with and without glaciers*

Based on our map analysis, the total stream length covered by PAs was 55 229 km, representing only 12% of the total length (Fig. 2a). Stream orders from 1 to 7 were covered by PAs, with lengths ranging from 2.5 to 4.0 km (Fig. 2a & b). We observed a high number of first-order streams, of which 14% were covered by PAs. Seventh-order streams were even less covered by PAs (6%; Fig. 2c). Thus, small streams such as headwaters and ephemeral streams seem to be better protected by PAs than larger streams and rivers, but the percentages remain low.

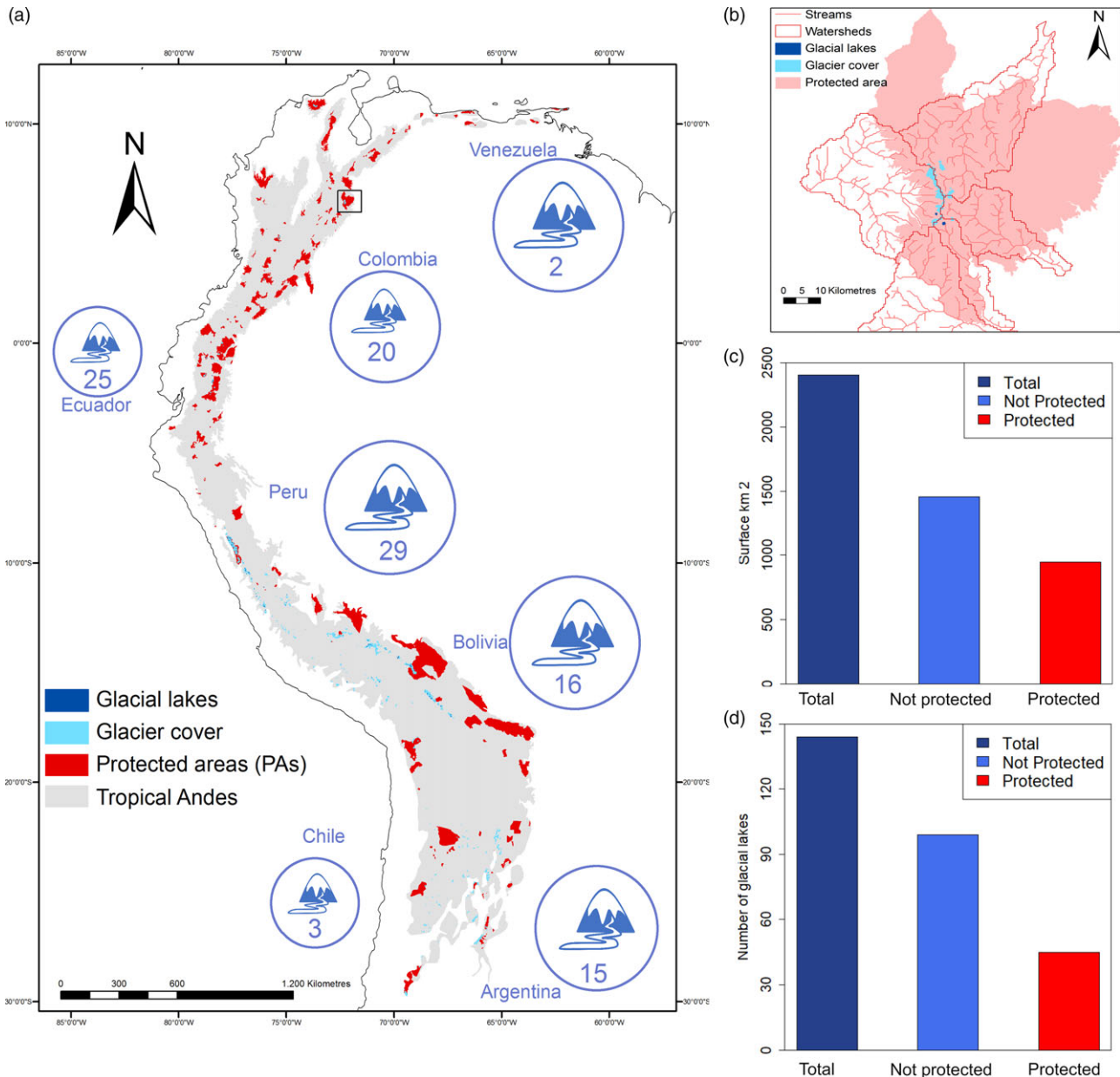
One study quantified the protection of streams and catchments by PAs (Thieme et al. 2007), including the longitudinal and lateral connectivity of drainage basins, ranging from high elevations in the Andes to lowland sites in the Amazon. PAs covered 10% (1259 km<sup>2</sup>) of the high-elevation streams within a buffer area of

10 km<sup>2</sup>, but without representation of longitudinal and lateral connectivity.

We did not find studies on glacier streams and catchments protected by PAs. In our maps, we observed that 40% of the total glacier surface area (2407 km<sup>2</sup>) was covered by PAs (Fig. 3c), which resulted in 120 watersheds with glaciers (average surface area of 631 km<sup>2</sup>) covered by PAs, representing 29% of watersheds with glaciers in the tropical Andes (Fig. 3a & b).

*Ramsar sites in the tropical Andes*

We found four research articles assessing the extent of and future changes to wetlands. Ramsar sites are increasing in number and surface area, resulting in 113 sites with a total area of c. 373 000 km<sup>2</sup> in South America, but more specific national policies and monitoring are needed (Wittmann et al. 2015).



**Fig. 3.** (a) Protected areas (PAs), glacial lakes, glacier cover and watershed number covered by glaciers and PAs in circles at the national scale (watershed average surface area = 631 km<sup>2</sup>). Only 110 watersheds covered with glaciers are shown as the remainder were shared among different countries. The PA within the black rectangle is the Cocuy National Park, which is depicted in (b). (c) Total surface area of PA protected and unprotected glacier cover. (d) Total number of PA protected and unprotected glacial lakes in the tropical Andes.

We quantified 31 Ramsar sites that occupy 60 232.6 km<sup>2</sup> (4%) of the tropical Andes (Table 1).

The *bofedales* (peatlands) in Peru were designated as a Ramsar site in 2003 (Ramsar Site No. 1317). In Ecuador, *bofedales* cover 205.18 km<sup>2</sup> (39%) of the total area of the Chimborazo Fauna Production Preserve, at 3800–6268 m altitude (Jara et al. 2019). In Peru, a total peatland area of 384.44 km<sup>2</sup> represents 11% of the area of the Huascarán National Park (Chimner et al. 2019). Moreover, Otto and Gibbons (2017) assessed peatland area responses under 2 precipitation scenarios in 17 watersheds located inside and outside of the Salinas y Agua Blanca National Reserve. Projections indicate that annual rainfall decrease will result in total wetland loss by the end of the twenty-first century for watersheds with 200–500 mm of annual mean rainfall and mainly inside of the

PA. In Sajama National Park of Bolivia, *bofedales* have declined in area from 34 km<sup>2</sup> in 1986 to 22 km<sup>2</sup> in 2016, whereas dry mixed grasses increased from 5.1 to 20.3 km<sup>2</sup> (Yager et al. 2019). Stakeholders also perceived a likely reduction in ecosystem services provision by peatlands (e.g., water storage) over the next 20 years in Salinas y Aguada Blanca National Reserve, Peru (Ibáñez Blancas et al. 2018).

#### *To what extent are ecosystem services and biodiversity safeguarded by current protection efforts?*

##### *PAs, freshwater biodiversity and impacts*

One of the most comprehensive reports on the distribution, extinction risk and climate change vulnerability of freshwater

**Table 1.** Ramsar sites and surface areas (km<sup>2</sup>) in the tropical Andes based on the Ramsar sites information databases (<https://rsis.ramsar.org>).

Name	Area (km <sup>2</sup> )	Country	Name	Area (km <sup>2</sup> )	Country	Name	Area (km <sup>2</sup> )	Country
Reserva Provincial Laguna Brava	4050.00	Argentina	Los Lipez (half protected by the National Andean Wildlife Reserve Eduardo Avaroa)	14 277.17	Bolivia	Parque Nacional Cajas	294.77	Ecuador
Laguna de los Pozuelos	162.24	Argentina	Lago Poopó y Uru	9676.07	Bolivia	Complejo Llanganati	303.55	Ecuador
Lagunas Altoandinas y Puneñas de Catamarca	12 281.75	Argentina	Cuenca de Tajzara	55.00	Bolivia	Complejo de Humedales de Núcachi Turupamba	122.90	Ecuador
Lagunas de Vilama	1570.00	Argentina	Lago Titicaca	12 600.00	Bolivia/Peru	Reserva Ecológica el Angel	170.03	Ecuador
Sistema Hidrológico de Soncor	671.33	Chile	Bofedales y Lagunas de Salinas	176.57	Peru	Laguna de la Cocha	390.00	Colombia
Salar de Huasco	60.00	Chile	Laguna del Indio y Dique de los Españoles	5.02	Peru	Complejo de Humedales del Alto Río Cauca Asociado a la Laguna de Sonso	55.25	Colombia
Salar de Surire	158.58	Chile	Reserva Nacional de Junin	530.00	Peru	Complejo de Humedales Laguna del Otún	65.79	Colombia
Complejo Lacustre Laguna del Negro Francisco y Laguna Santa Rosa	624.60	Chile	Humedal Lucre-Huacarpay	19.79	Peru	Complejo de Humedales del Distrito Capital de Bogotá	6.67	Colombia
Salar de Tara	964.39	Chile	Lagunas las Arreviatadas	12.50	Peru	Sistema Lacustre de Chingaza	40.58	Colombia
Salar de Pujsa	173.97	Chile	Sistema Lacustre Yacuri	277.62	Ecuador			
Aguas Calientes IV	155.29	Chile	Lagunas de Compadre	281.15	Ecuador			

biodiversity in the tropical Andes and Amazon is that of Tognelli et al. (2016). It includes 9 freshwater ecoregions, from Bolivia to Colombia, and 967 species in 4 taxonomic groups: fishes, molluscs, dragonflies and aquatic plants. The study identified 86 key biodiversity areas within 22 PAs and 39 catchment management zones, and it proposed 25 new freshwater PAs to include 151 threatened species. Furthermore, Tognelli et al. (2019) reported that 571 endemic fish species (88%) were not within any PA, and they identified c. 475 catchments as very high priority for biodiversity conservation, but only 2% were covered by PAs. Furthermore, Miranda et al. (2018) reported fish distribution in clusters at the Suaza River (Colombia), with a clear distinction in the communities at the headwater sites (within the Cueva de los Guácharos National Park) due to the low level of human intervention. Based on the distribution of 481 vertebrates (including fishes) and 54 invertebrates across elevational gradients in an Ecuadorian Andean–Amazon basin and covering 6 PAs, Lessmann et al. (2016) reported that lowlands (<600 m) are diverse in vertebrate species, whereas mid-elevations (600–1600 m) are diverse in invertebrates, suggesting the need for fluvial corridors for sustaining species migration. On the other hand, an analysis of the conservation of 132 aquatic macroinvertebrate species within PAs from the tropical Andes to the dry Chaco reported that only 0.009% of the macroinvertebrate distribution was within the PAs (Nieto et al. 2017).

Regarding climate-related impacts, Tognelli et al. (2016) reported aquatic plants as the most vulnerable to climate change, followed by fishes, under the IPCC scenario RCP 4.5. Climate change will likely reduce the dispersion of most of the Andean–Amazon fish species, especially those inhabiting the highlands due to the additional effects of fragmentation (Herrera et al. 2020). Currently, 142 dams have impacted the connectivity of six major Andean–Amazon river basins, and future new dams will likely impact the connectivity of 5 basins and 671 fish species (Anderson et al. 2018). On the other hand, lakes in Cajas National Park (Ecuador) have started to become thermally stratified, resulting in plankton community shifts and with likely impacts on higher trophic levels such as fishes (Labaj et al. 2017, 2018a, 2018b).

### PA's and ecosystem services

#### Fisheries and organic matter decomposition

One study assessed a PA role in organic matter decomposition by aquatic invertebrates, where leaf litter decomposition inside of Cajas National Park (Ecuador) was three times faster than outside of the PA (Rincon et al. 2017). One study on fishery resources in Ramsar sites of the Andes reported a biomass production of 49 631 tonnes by the year 2000 of the native species *Orestias ispi* (Vila et al. 2007).

#### Carbon fluxes

Carbon flux studies have been conducted mainly in *bofedales*, which are carbon dioxide sinks because organic matter production is greater than decomposition (Hribljan et al. 2016). Carbon stocks of 1 040 000 and 572 000 kg ha<sup>-1</sup> were reported for the *bofedales* of Sajama National Park and Tuni in Bolivia, respectively (Hribljan et al. 2015). Across hydrological gradients, undrained *bofedales* within the Huascanan National Park in Peru are likely to be carbon sinks, and highly drained peatlands are carbon sources (Planas-Clarke et al. 2020).



### Water supply

In the grey literature, we found that the city of Tarija (Bolivia) gets 70% and 30% of its surface water and groundwater, respectively, from the La Vitoria watershed, which is protected by Cordillera de Sama National Park (Brown 2005). Other studies have assessed the water supply by forests in PAs (Ramos Franco & Armenteras Pascual 2019, Fastré et al. 2020). However, we did not find studies that assessed the direct role of PAs in water supply.

## Discussion

### Knowledge gaps and future studies

Our review has focused on PAs because they are key instruments in the creation of environmental policies for ecosystem conservation and climate change adaptation. We synthesized evidence of the role of PAs in freshwater ecosystem conservation in the tropical Andes, including ecosystems at very high elevation. Addressing the following knowledge gaps would contribute to freshwater conservation and climate change adaptation.

### The creation of PAs

PAs have not been designed for the protection of large rivers and watersheds, and they mainly protect small freshwater ecosystems (e.g., small headwater catchments and streams). Furthermore, the creation of local PAs highlights the need to maintain the water supply. Therefore, it is important to know more about the creation of specific PAs for freshwater conservation, which may have been based on very local decisions (e.g., by rural communities and private sectors; Hora et al. 2018). Mapping these local PAs and other adaptation measures that explore nature-based solutions, including the protection and restoration of freshwater ecosystems to prevent climate-related risks such as hydrological changes (Hartman et al. 2016), could provide a well-rounded assessment of current climate change adaptation and water provision efforts. Additionally, more effective efforts are needed for PAs without management plans and with high values for glacier protection.

### Freshwater ecosystems covered by PAs

To our knowledge, this review is the first to combine information about freshwater systems and PAs, including glaciated systems, in the tropical Andes. Our mapping quantified lower percentages of lakes, glacier lakes, streams and glaciers covered by PAs than the 17% of the Aichi protection target (Convention on Biological Diversity 2010). However, we might have underestimated or overestimated the real conditions as we used global inventories and because freshwater ecosystems and glaciers are dynamic and changing in response to climatic conditions (Baraer et al. 2012, Vuille et al. 2018). Therefore, more detailed inventories within and around PAs are needed for the better design of freshwater conservation. Future lake inventories might include climate-related impacts such as the increase in the number of lakes due to glacier retreat (Drenkhan et al. 2018) and water evaporation (Sotgiu et al. 2017). However, PAs alone or the increase in the area of PAs might not be sufficient to guarantee freshwater ecosystem conservation and to cope with climate change as in terrestrial ecosystems (Possingham et al. 2015, Pringle et al. 2017). The achievement of long-term freshwater conservation and climate change adaptation requires strong cooperation among stakeholders and local and regional governments in order to put in place

effective and sustainable management measures of freshwater ecosystems within PAs.

### PAs, aquatic biodiversity and ecosystem services

Biodiversity conservation within PAs has been assessed mainly for fishes across the Andean–Amazon region, while few studies have been performed on the western Andean slope and at high-elevation sites (>4000 m altitude). We have provided a benchmark of the current situation, but more research is needed for the habitat conservation planning of other aquatic groups (e.g., benthic invertebrates; Crespo-Pérez et al. 2020) and for supporting species' upward migration in mountains due to increasing warming or the new habitats that have been opened up by glacier retreat (Milner et al. 2008, Seimon et al. 2017). Regarding river connectivity, future studies on the protection of free-flowing rivers by PAs across the Andean–Amazon region could provide relevant information about fluvial connectivity, conservation of river ecological function and ecosystem services provision (Opperman et al. 2021). Furthermore, we evidenced how glacier coverages have been protected by PAs as they represent important water sources in the region. Nevertheless, polar and alpine sites that include ice sheets, glaciers and perennial snowfields have been categorized as a new biome by the IUCN due to their high concentrations of microbial life (Keith et al. 2020). Investigations of the biodiversity inhabiting glaciers are needed in order to better understand their implications for stream and lake functioning in the tropical Andes, as has been reported for other regions (Fellman et al. 2015).

On the other hand, studies of PAs as constant providers of ecosystem services were limited to specific cases of fisheries and organic matter decomposition. Studies on the role of organic matter decomposition by aquatic invertebrates for maintaining water quality inside and outside PAs are required. Furthermore, we did not find any studies about the economic implications of water supply by PAs. Specifically, we did not find studies that have assessed the economic value of PAs in protecting significant water sources such as glaciers, which are of great importance for the millions of people inhabiting the region (Soruco et al. 2015, Kinouchi et al. 2019). Thus, future studies focusing on these issues will provide relevant information for freshwater conservation in the tropical Andes.

**Supplementary material.** To view supplementary material for this article, please visit <https://doi.org/10.1017/S0376892921000382>.

**Acknowledgements.** We thank the anonymous reviewer and editor for their helpful suggestions and corrections regarding a previous version of the manuscript. We thank Gabrielle Vance from the Mountain Research Initiative (MRI) for the English revision.

**Financial support.** Writing this review article was an activity of the Mentoring and Training Program in IPCC Processes for Early Career Mountain Researchers supported by the Swiss Agency for Development and Cooperation (SDC) within the programme of SDC contributions to the 6th Assessment Cycle of the IPCC, and organized by MRI, University of Zurich, Helvetas and ICIMOD. E. Quenta-Herrera conceptualized the review article and was selected to participate in the programme.

**Conflict of interest.** None.

**Ethical standards.** None.



## References

- Abell R, Lehner B, Thieme M, Linke S (2017) Looking beyond the fence line: assessing protection gaps for the world's rivers. *Conservation Letters* 10: 384–394.
- Acreman M, Hughes KA, Arthington AH, Tickner D, Dueñas MA (2020) Protected areas and freshwater biodiversity: a novel systematic review distills eight lessons for effective conservation. *Conservation Letters* 13: e12684.
- Anderson EP, Jenkins CN, Heilpern S, Maldonado-Ocampo JA, Carvajal-Vallejos FM, Encalada AC et al. (2018) Fragmentation of Andes-to-Amazon connectivity by hydropower dams. *Science Advances* 4: ea01642.
- Azevedo-Santos VM, Frederico RG, Fagundes CK, Pompeu PS, Pelicice FM, Padial AA et al. (2019) Protected areas: a focus on Brazilian freshwater biodiversity. *Diversity and Distributions* 25: 442–448.
- Baraer M, Mark B, McKenzie J, Condom T, Bury J, Huh K et al. (2012). Glacier recession and water resources in Peru's Cordillera Blanca. *Journal of Glaciology* 58: 134–150.
- Biggs J, von Fumetti S, Kelly-Quinn M (2017) The importance of small waterbodies for biodiversity and ecosystem services: implications for policy makers. *Hydrobiologia* 793: 3–39.
- Bradley RS, Vuille M, Diaz HF, Vergara W (2006) Threats to water supplies in the tropical Andes. *Science* 312: 1755–1756.
- Brown M (2005) Case study of watershed valuation in the Sama Biological Reserve, Bolivia. *The Nature Conservancy* [www document]. URL <https://www.conservationgateway.org/Documents/Krchnak%202007.pdf>
- Chimner RA, Bourgeau-Chavez L, Grelik S, Hribljan JA, Clarke AM, Polk MH et al. (2019) Mapping mountain peatlands and wet meadows using multi-date, multi-sensor remote sensing in the Cordillera Blanca, Peru. *Wetlands* 39: 1057–1067.
- Colonia D, Torres J, Haerberli W, Schauwecker S, Braendle E, Giraldez C, Cochachin A (2017) Compiling an inventory of glacier-bed overdeepenings and potential new lakes in de-glaciating areas of the Peruvian Andes: approach, first results, and perspectives for adaptation to climate change. *Water* 9: 336.
- Constantine JA, Dunne T, Ahmed J, Legleiter C, Lazarus ED (2014) Sediment supply as a driver of river meandering and floodplain evolution in the Amazon Basin. *Nature Geoscience* 7: 899–903.
- Convention on Biological Diversity (2010) Aichi Biodiversity Targets [www document]. URL <https://www.cbd.int/sp/targets/>
- Cook SJ, Kougkoulos I, Edwards LA, Dortch J, Hoffmann D (2016) Glacier change and glacial lake outburst flood risk in the Bolivian Andes. *Cryosphere* 10: 2399–2413.
- Crespo-Pérez V, Dangles O, Ibarra C, Espinosa R, Andino P, Jacobsen D, Cauvy-Fraunié S (2020) Functional structure and diversity of invertebrate communities in a glacierised catchment of the tropical Andes. *Freshwater Biology* 65: 1348–1362.
- Drenkhan F, Guardamino L, Huggel C, Frey H (2018) Current and future glacier and lake assessment in the deglaciating Vilcanota–Urubamba basin, Peruvian Andes. *Global and Planetary Change* 169: 105–118.
- Dudley N, Stolton S, Belokurov A, Krueger L, Lopoukhine N, MacKinnon K et al. (2010) Natural solutions: protected areas helping people cope with climate change. IUCN/WWF, TNC, UNDP, WCS, The World Bank and WWF [www document]. URL <https://portals.iucn.org/library/node/9433>
- Dudley N, Shadie P, Stolton S (2013) Guidelines for applying protected area management categories including IUCN WCPA best practice guidance on Recognizing Protected Areas and Assigning Management Categories and Governance. *Best Practice Protected Area Guidelines Series* [www document]. URL [http://www.iucn.org/pa\\_categories](http://www.iucn.org/pa_categories)
- Elsen PR, Monahan WB, Merenlender AM (2018) Global patterns of protection of elevational gradients in mountain ranges. *Proceedings of the National Academy of Sciences of the United States of America* 115: 6004–6009.
- Fastré C, Possingham HP, Strubbe D, Matthysen E (2020) Identifying trade-offs between biodiversity conservation and ecosystem services delivery for land-use decisions. *Scientific Reports* 10: 7971.
- Fellman JB, Hood E, Raymond PA, Hudson J, Bozeman M, Arimitsu M (2015) Evidence for the assimilation of ancient glacier organic carbon in a proglacial stream food web. *Limnology and Oceanography* 60: 1118–1128.
- GLIMS, NSIDC (2018) Global Land Ice Measurements from Space glacier database. Compiled and made available by the international GLIMS community and the National Snow and Ice Data Center, Boulder, CO, USA [www document]. URL <https://doi.org/10.7265/N5V98602>
- Hansen WF (2001) Identifying stream types and management implications. *Forest Ecology and Management* 143: 39–46.
- Hartman BD, Bookhagen B, Chadwick OA (2016), The effects of check dams and other erosion control structures on the restoration of Andean *bofedal* ecosystems. *Restoration Ecology* 24: 761–772.
- Herrera RGA, Oberdorff T, Anderson EP, Brosse S, Carvajal-Vallejos FM, Frederico RG et al. (2020) The combined effects of climate change and river fragmentation on the distribution of Andean Amazon fishes. *Global Change Biology* 26: 5509–5523.
- Hoffman M, Koenig K, Bunting G, Costanza J, Williams KJ (2016) Biodiversity hotspots (version 2016.1) [data set]. Zenodo [www document]. URL <http://doi.org/10.5281/zenodo.3261807>
- Hoffmann D, Oetting I (2011) Climate change and protected areas in Bolivia. In: *The Economic, Social and Political Elements of Climate Change*, ed. W Leal Filho (pp. 355–372). Berlin, Germany: Springer.
- Hora B, Marchant C, Borsdorf A (2018) Private protected areas in Latin America: between conservation, sustainability goals and economic interests. a review. *Journal on Protected Mountain Areas Research and Management* 10: 87–94.
- Hribljan, JA, Cooper DJ, Sultenfuss J, Wolf EC, Heckman KA, Lilleskov EA, Chimner RA (2015) Carbon storage and long-term rate of accumulation in high-altitude Andean peatlands of Bolivia. *Mires and Peat* 15: 1–14.
- Hribljan JA, Suárez E, Heckman KA, Lilleskov EA, Chimner RA (2016) Peatland carbon stocks and accumulation rates in the Ecuadorian páramo. *Wetlands Ecology and Management* 24: 113–127.
- Ibáñez Blancas AN, La Torre-Cuadros M, Mallma Carrera GA (2018) Using foresight to gain a local perspective on the future of ecosystem services in a mountain protected area in Peru. *Mountain Research and Development* 38: 192–202.
- Íñiguez Gallardo MV, Helsley J, Pinel S, Ammon J, Lopez Rodriguez FV, Wendland K (2013) Collaborative community-based governance in a trans-boundary wetland system in the Ecuadorian Andes: opportunities and challenges at a proposed Ramsar site. *Mountain Research and Development* 33: 269–279.
- Jantz P, Goetz S, Laporte N (2014) Carbon stock corridors to mitigate climate change and promote biodiversity in the tropics. *Nature Climate Change* 4: 138–142.
- Jara C, Delegido J, Ayala J, Lozano P, Armas A, Flores V (2019) Study of wetlands in the Ecuadorian Andes through the comparison of Landsat-8 and Sentinel-2 images. *Revista de Teledetección* 53: 45–57.
- Jiménez Cisneros BE, Oki T, Arnell NW, Benito G, Cogley JG, Döll P et al. (2014) Freshwater resources. In: *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, ed. CB Field, VR Barros, DJ Dokken, KJ Mach, MD Mastrandrea, TE Bilir et al. (pp. 229–269). Cambridge, UK: Cambridge University Press.
- Joslin A (2020) Translating water fund payments for ecosystem services in the Ecuadorian Andes. *Development and Change* 51: 94–116.
- Keith D, Ferrer-Paris J, Nicholson E, Kingsford R (2020) *The IUCN Global Ecosystem Typology 2.0: Descriptive Profiles for Biomes and Ecosystem Functional Groups*. Gland, Switzerland: IUCN.
- Kinouchi T, Nakajima T, Mendoza J, Fuchs P, Asaoka Y (2019) Water security in high mountain cities of the Andes under a growing population and climate change: a case study of La Paz and El Alto, Bolivia. *Water Security* 6: 100025.
- Labaj AL, Michelutti N, Smol JP (2017) Changes in cladoceran assemblages from tropical high mountain lakes during periods of recent climate change. *Journal of Plankton Research* 39: 211–219.
- Labaj AL, Michelutti N, Smol JP (2018a) Annual stratification patterns in tropical mountain lakes reflect altered thermal regimes in response to climate change. *Fundamental and Applied Limnology* 191: 267–275.
- Labaj AL, Michelutti N, Smol JP (2018b) Cladocera in shallow lakes from the Ecuadorian Andes show little response to recent climate change. *Hydrobiologia* 822: 203–216.

- Lehner B, Verdin K, Jarvis A (2008) New global hydrography derived from spaceborne elevation data. *Eos* 89: 93–94.
- Leroy D (2019) Farmers' perceptions of and adaptations to water scarcity in Colombian and Venezuelan *paramos* in the context of climate change. *Mountain Research and Development* 39: R21–R34.
- Lessmann J, Guayasamin JM, Casner KL, Flecker AS, Funk WC, Ghalambor CK et al. (2016) Freshwater vertebrate and invertebrate diversity patterns in an Andean–Amazon basin: implications for conservation efforts. *Neotropical Biodiversity* 2: 99–114.
- McDowell G, Huggel C, Frey H, Wang FM, Cramer K, Ricciardi V (2019) Adaptation action and research in glaciated mountain systems: are they enough to meet the challenge of climate change? *Global Environmental Change* 54: 19–30.
- Message ML, Lehner B, Grill G, Nedeva I, Schmitt O (2016) Estimating the volume and age of water stored in global lakes using a geo-statistical approach. *Nature Communications* 7: 13603.
- Messerli B, Grosjean M, Vuille M (1997) Water availability, protected areas, and natural resources in the Andean desert Altiplano. *Mountain Research and Development* 17: 229–238.
- Milner AM, Robertson AL, Monaghan KA, Veal AJ, Flory EA (2008). Colonization and development of an Alaskan stream community over 28 years. *Frontiers in Ecology and the Environment* 6: 413–419.
- Miranda R, Tobes I, Gaspar S, Peláez-Rodríguez M (2018) Patterns in the distribution of fish assemblages and their association with habitat variables in the Suaza River on its way through the Cueva de los Guácharos National Park, Colombia. *Ecoscience* 25: 85–95.
- Moher D, Liberati A, Tetzlaff J, Altman DG (2009) Preferred reporting items for systematic reviews and meta-analyses: the PRISMA statement. *PLoS Medicine* 6: e1000097.
- Montoya-Zumaeta J, Rojas E, Wunder S (2019) Adding rewards to regulation: the impacts of watershed conservation on land cover and household wellbeing in Moyobamba, Peru. *PLoS ONE* 14: e0225367.
- Mosquera PV, Hampel H, Vázquez RF, Alonso M, Catalan J (2017) Abundance and morphometry changes across the high-mountain lake-size gradient in the tropical Andes of southern Ecuador. *Water Resources Research* 53: 7269–7280.
- Myers N, Mittermeier R, Mittermeier C, da Fonseca G, Kent J (2000) Biodiversity hotspots for conservation priorities. *Nature* 403: 853–858.
- Nieto C, Ovando XM, Loyola R, Izquierdo A, Romero F, Molineri C et al. (2017) The role of macroinvertebrates for conservation of freshwater systems. *Ecology and Evolution* 7: 5502–5513.
- Opperman JJ, Shahbol N, Maynard J, Grill G, Higgins J, Tracey D, Thieme M (2021) Safeguarding free-flowing rivers: the global extent of free-flowing rivers in protected areas. *Sustainability* 13: 2805.
- Ostovar AL (2019) Investing upstream: watershed protection in Piura, Peru. *Environmental Science & Policy* 96: 9–17.
- Otto M, Gibbons RE (2017) Potential effects of projected decrease in annual rainfall on spatial distribution of high Andean wetlands in southern Peru. *Wetlands* 37: 647–659.
- Pepin N, Bradley RS, Diaz HF, Baraer M, Caceres EB, Forsythe N et al. (2015) Elevation-dependent warming in mountain regions of the world. *Nature Climate Change* 5: 424–430.
- Pillco Zolá R, Bengtsson L, Berndtsson R, Marti-Cardona B, Satgé F, Timouk F et al. (2019) Modelling Lake Titicaca's daily and monthly evaporation. *Hydrology and Earth System Sciences* 23: 657–668.
- Planas-Clarke AM, Chimner RA, Hribljan JA, Lilleskov EA, Fuentealba B (2020) The effect of water table levels and short-term ditch restoration on mountain peatland carbon cycling in the Cordillera Blanca, Peru. *Wetlands Ecology and Management* 28: 51–69.
- Polk MH, Young KR, Baraer M, Mark BG, McKenzie JM, Bury J, Carey M (2017) Exploring hydrologic connections between tropical mountain wetlands and glacier recession in Peru's Cordillera Blanca. *Applied Geography* 78: 94–103.
- Possingham, HP, Bode M, Klein C (2015) Optimal conservation outcomes require both restoration and protection. *PLoS Biology* 13: e1002052.
- Pringle R (2017) Upgrading protected areas to conserve wild biodiversity. *Nature* 546: 91–99.
- Ramos Franco A, Armenteras Pascual D (2019) Interception and runoff of high Andean forest in the protective forest reserve 'El Malmo'. *Acta Biológica Colombiana* 24: 97–108.
- Ramsar (2020) Ramsar Sites Information Service [www document]. URL <https://rsis.ramsar.org/>
- Redondo-Brenes A (2009) An assessment of a fund for watershed conservation in the southern Ecuadorian Andes. *Journal of Sustainable Forestry* 28: 755–776.
- Rincon J, Merchan D, Sparer A, Rojas D, Zarate E (2017) Leaf-litter breakdown as a tool to assess the functional integrity of high Andean streams of Southern Ecuador. *Revista de Biología Tropical* 65: 321–334.
- Rundel P, Palma B (2000) Preserving the unique *puna* ecosystems of the Andean Altiplano: a descriptive account of Lauca National Park, Chile. *Mountain Research and Development* 20: 262–271.
- Satgé F, Espinoza R, Zolá RP, Roig H, Timouk F, Molina J et al. (2017) Role of climate variability and human activity on Poopó Lake droughts between 1990 and 2015 assessed using remote sensing data. *Remote Sensing* 9: 218.
- Seimon TA, Seimon A, Yager K, Reider K, Delgado A, Sowell P et al. (2017) Long-term monitoring of tropical alpine habitat change, Andean anurans, and chytrid fungus in the Cordillera Vilcanota, Peru: results from a decade of study. *Ecology and Evolution* 7: 1527–1540.
- Shugar DH, Burr A, Haritashya UK, Kargel JS, Watson CS, Kennedy MC et al. (2020) Rapid worldwide growth of glacial lakes since 1990. *Nature Climate Change* 10: 939–945.
- Soruco A, Vincent C, Rabatel A, Francou B, Thibert E, Sicart JE, Condom T (2015) Contribution of glacier runoff to water resources of La Paz city, Bolivia (16 degrees S). *Annals of Glaciology* 56: 147–154.
- Thieme M, Lehner B, Abell R, Hamilton SK, Kellendorfer J, Powell G, Riveros JC (2007) Freshwater conservation planning in data-poor areas: an example from a remote Amazonian basin (Madre de Dios River, Peru and Bolivia). *Biological Conservation* 135: 484–501.
- Tognelli MF, Anderson EP, Jiménez-Segura LF, Chuctay J, Chocano L, Maldonado-Ocampo JA et al. (2019) Assessing conservation priorities of endemic freshwater fishes in the tropical Andes region. *Aquatic Conservation: Marine and Freshwater Ecosystems* 29: 1123–1132.
- Tognelli MF, Lasso CA, Bota-Sierra CA, Jiménez-Segura LF, Cox NA (2016) *Estado de Conservación y Distribución de la Biodiversidad de Agua Dulce en los Andes Tropicales*. Gland, Switzerland: IUCN.
- Vila I, Pardo R, Scott S (2007) Freshwater fishes of the Altiplano. *Aquatic Ecosystem Health & Management* 10: 201–211.
- Vuille M, Carey M, Huggel C, Buytaert W, Rabatel A, Jacobsen D et al. (2018) Rapid decline of snow and ice in the tropical Andes – impacts, uncertainties and challenges ahead. *Earth-Science Reviews* 176: 195–213.
- Walker R, Moore NJ, Arima E, Perz S, Simmons C, Caldas M et al. (2009) Protecting the Amazon with protected areas. *Proceedings of the National Academy of Sciences of the United States of America* 106: 10582–10586.
- Wittmann F, Householder E, Wittmann A, Lopes A, Junk WJ, Piedade MT (2015) Implementation of the Ramsar Convention on South American wetlands: an update. *Research and Reports in Biodiversity Studies* 4: 47–58.
- Yager K, Valdivia C, Slayback D, Jimenez E, Isela Meneses R, Palabral A et al. (2019) Socio-ecological dimensions of Andean pastoral landscape change: bridging traditional ecological knowledge and satellite image analysis in Sajama National Park, Bolivia. *Regional Environmental Change* 19: 1353–1369.
- Young KR (1993) National park protection in relation to the ecological zonation of a neighboring human community: an example from northern Peru. *Mountain Research & Development* 13: 267–280.
- Zalazar L, Ferri L, Castro M, Gargantini H, Gimenez M, Pitte P et al. (2020) Spatial distribution and characteristics of Andean ice masses in Argentina: results from the first National Glacier Inventory. *Journal of Glaciology* 66: 938–949.
- Zimmer A, Meneses RI, Rabatel A, Soruco A, Dangles O, Anthelme F (2018) Time lag between glacial retreat and upward migration alters tropical alpine communities. *Perspectives in Plant Ecology, Evolution and Systematics* 30: 89–102.